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# Soil biotic legacy effects of extreme weather events influence plant invasiveness

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Climate change is expected to increase future abiotic stresses on ecosystems through extreme weather events leading to more extreme drought and rainfall incidences [Jentsch A, et al. (2007) *Front Ecol Environ* 5(7):365–374]. These fluctuations in precipitation may affect soil biota, soil processes [Evans ST, Wallenstein MD (2012) *Biogeochemistry* 109:101–116], and the proportion of exotics in invaded plant communities [Jiménez MA, et al. (2011) *Ecol Lett* 14:1277–1235]. However, little is known about legacy effects in soil on the performance of exotics and natives in invaded plant communities. Here we report that drought and rainfall effects on soil processes and biota affect the performance of exotics and natives in plant communities. We performed two mesocosm experiments. In the first experiment, soil without plants was exposed to drought and/or rainfall, which affected soil N availability. Then the initial soil moisture conditions were restored, and a mixed community of co-occurring natives and exotics was planted and exposed to drought during growth. A single stress before or during growth decreased the biomass of natives, but did not affect exotics. A second drought stress during plant growth reset the exotic advantage, whereas native biomass was not further reduced. In the second experiment, soil inoculation revealed that drought and/or rainfall influenced soil biotic legacies, which promoted exotics but suppressed natives. Our results demonstrate that extreme weather events can cause legacy effects in soil biota, promoting exotics and suppressing natives in invaded plant communities, depending on the type, frequency, and timing of extreme events.

invaded ecosystems | nitrogen cycle | plant invasion | plant–soil interaction | soil microbes

Extreme weather events are expected to increase future abiotic stresses on ecosystems, but the outcomes of these events on the performance of exotic and native plant species in plant communities and the involved mechanisms are poorly understood (1–3). Patterns of drought and rainfall have been predicted to become more extreme under climate change (4, 5). Intensive drought and rainfall can directly affect the productivity and composition of plant communities through altered water supply (6, 7). This may lead to altered dominance of exotic plant species in invaded ecosystems (3, 8–10) when exotics and natives respond differently to extreme weather events (11). Whether extreme weather events also might change plant community composition through changes in soil conditions is less clear. Variations in climate are known to influence the composition of soil microbial communities (12), but the legacy effects in soil of such changes on exotics and natives have not been investigated. In the present study, we examined how extreme weather event-induced changes in soil processes and biota may influence the biomass response of exotics and natives in mixed plant communities. In two mesocosm experiments, we tested how soil biota and the processes that they drive may explain the responses of exotic and native plant species to single and repeated extreme weather events. This knowledge is important for evaluating and improving predictions of current and future climate change on the composition of plant communities.

Extreme drought and rainfall are known to influence soil biota directly (13–16). Drought generally reduces microbial activities

in soil, and on rewetting a short-term increase in microbial activity occurs (17–19). Apart from this activity, the microbial community composition also can be altered by exposing soil to drought and rewetting (14, 18, 20). Moreover, previous experiments on drought and rainfall events have shown that such treatments remain as a legacy in soil microbial communities, which become apparent when testing their responses to an additional drying-rewetting event (21). However, whether and how such legacy effects of extreme weather events on soil biota may affect plant species composition is unknown (22).

Extreme weather events may promote exotics over native plant species through several mechanisms. Exotics that become dominant in native vegetation often have specific traits associated with better performance than natives, such as faster growth rates, which may enable them to recover faster after a pulse of abiotic stress (23). Another possibility is that extreme weather events indirectly influence exotics and natives in plant communities via soil biota (24), which can influence interactions between plant species (22, 25). Exotics may be less influenced than natives by effects of extreme weather events on soil biota, because exotics experience weaker negative and positive effects from interactions with soil biota (26–28). For example, promotion of soil pathogens by extreme weather events may reduce the proportion of natives, because exotics are less influenced by soil pathogens in the new range (27, 29). On the other hand, when exotic plant species lack a co-evolutionary history with the symbionts, positive effects of plant growth-promoting soil biota might enable native plant species to cope better with drought stresses (30). Thus, if soil biota effects on plants are enforced by extreme weather events either before or during plant growth, then exotic and native plant species might be facilitated or suppressed differently by soil biotic legacy effects.

We tested the hypothesis that a history of a drought and/or rainfall event before plant growth would affect exotic and native plant biomass in mixed plant communities via changed soil conditions. We studied how natives and exotics respond to an additional drought event during plant growth when grown in soils with or without a legacy of drought and/or rainfall. We tested our hypothesis in a mesocosm experiment, exposing the mesocosms to eight different drought and rainfall scenarios (Fig. S1). These are realistic scenarios predicted for northwestern Europe (4). In experiment 1, the mesocosms were filled with field soil and then exposed to an extreme drought and/or rainfall event before planting. Then the original soil moisture conditions were restored, and a mixed plant

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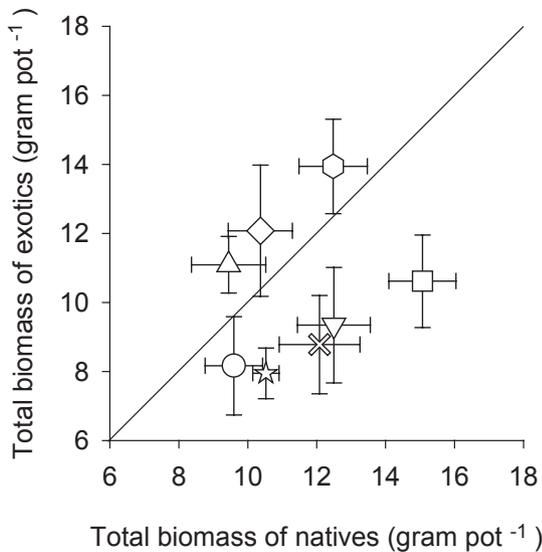
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**Fig. 1.** Exotic versus native species responses to water stresses. Biomass (mean  $\pm$  SE) of exotics and natives as affected by the different water stresses. Plant communities were exposed to control conditions (square;  $n = 7$ ), rainfall before plant growth (down-pointing triangle;  $n = 7$ ), drought during plant growth (up-pointing triangle;  $n = 7$ ), drought and rainfall before plant growth (diamond;  $n = 7$ ), drought before and during plant growth (circle;  $n = 8$ ), or rainfall before and during plant growth (star;  $n = 7$ ). The 1:1 line indicates when exotics and natives have the same biomass.

community of exotic and related (congeneric) native species was planted (Table S1). All plant species used in our study co-occur in invaded riverine ecosystems of northwestern Europe. Soil processes, microbial biomass, and fungal biomass were measured before planting. During plant growth, one-half of the mesocosms were exposed to an extreme drought event (Materials and Methods and Fig. S1).

In a second mesocosm experiment, we examined whether the shifts in the plant community biomass between natives and exotics observed in the first experiment might have been related to legacy effects of extreme weather events on soil biota (including symbionts and pathogens) (31). We tested this by inoculating living or sterilized soil from the first experiment in sterilized field soil (1:8 proportion; the sterilized field soil was of the same origin as that used in experiment 1) and growing the same plant communities again under the same conditions as in experiment 1.

## Results

The results support our hypothesis that extreme weather events before the growth season leave legacies that affect the performance of exotics and natives during the growth season. The drought and rainfall treatments influenced the biomass of both exotics ( $F_{7,49} = 2.33$ ,  $P = 0.04$ ) and natives ( $F_{7,49} = 3.67$ ,  $P = 0.003$ ). Drought and/or rainfall reduced the biomass of the native plant species in most treatments (Fig. 1 and Fig. S2B). In contrast, exotics were resistant to a single drought and/or rainfall before or during plant growth (Fig. 1 and Fig. S2A). However, in soil with a legacy of drought, exotics produced less biomass when exposed to drought during plant growth compared with their performance in soil exposed to drought before plant growth (Fig. 1 and Fig. S2A). In contrast, natives did not produce less biomass when exposed to drought during plant growth compared with their performance in soil preexposed to drought (Fig. 1 and Fig. S2B). Redundancy analysis revealed that four of the five exotics responded positively and all natives responded negatively

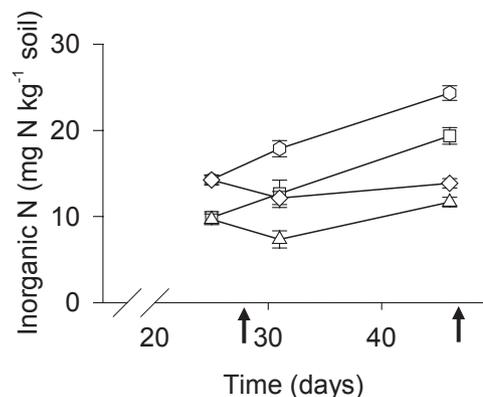
to the legacy of drought before plant growth (Fig. S3). In contrast, drought before and during plant growth resulted in the opposite response (Fig. S3). Drought before plant growth tended to increase the average ratio of exotic to total plant pair biomass (Table S2). On average, drought during plant growth reduced the biomass of the total plant community from  $24.0 \pm 0.6$  g to  $19.3 \pm 0.5$  g ( $F_{7,49} = 6.31$ ,  $P < 0.001$ ) (Fig. S2C).

The potential N mineralization (arginine-N mineralization) rate was influenced by drought and rainfall before plant growth, but the direction of the effects depended on the type of extreme event (Table S3 and Fig. S4). Drought increased and rainfall decreased N mineralization before soil moisture was reset. By the time soil moisture was restored and before plant growth, the N mineralization rate was still marginally lower in soil preexposed to both drought and rainfall compared with control soils. Soil respiration was not affected by drought and/or rainfall (Table S3). Rainfall before plant growth slightly decreased total microbial biomass (Fig. S5). Inorganic N in control soil remained different from that in soil exposed to drought and/or rainfall before plant growth. The availability of inorganic N immediately before planting was highest in soil exposed to drought and lowest in soil exposed to rainfall (Fig. 2 and Table S3). In mesocosms not exposed to additional drought during plant growth, total plant biomass was higher in soil with higher N availability (Spearman rho = 1,  $P < 0.01$ ;  $n = 4$ ). In these mesocosms, exotic species appeared to be more favored than natives by the higher N concentrations (Fig. S6).

In the inoculation experiment, drought and rainfall before plant growth influenced the effect of soil biota on exotics ( $F_{3,24} = 3.75$ ,  $P = 0.024$ ) and natives ( $F_{3,24} = 4.49$ ,  $P = 0.011$ ). Interestingly, soil biota exposed to drought and/or rainfall before plant growth had a positive effect on exotics and a negative effect on natives, whereas soil biota from control soil had the opposite effect (Fig. 3). In contrast to the responses of the exotics and natives, the soil biotic effect on the total plant community was unaffected by a legacy of drought and/or rainfall ( $F_{3,24} = 0.53$ ,  $P = 0.67$ ), demonstrating that native biomass had been replaced by exotic biomass.

## Discussion

Our mesocosm study reveals a mechanism by which extreme weather events influence the performance of exotics and natives in plant communities, namely through legacy effects of drought and



**Fig. 2.** Extreme event effects on soil N concentration over time. Mean  $\pm$  SE values are presented for control soil (square;  $n = 8$ ), soil with legacies of drought before plant growth (hexagon;  $n = 8$ ), rainfall before plant growth (triangle;  $n = 8$ ), or soils with both legacies (diamond;  $n = 8$ ). The first arrow at 28 d indicates up to when one-half of the soils were exposed to drought. At day 28, one-half of the mesocosms received rainfall before plant growth. The second arrow at day 48 indicates when plant communities were established. The first sampling point was during the drought stress, the second sampling point was 3 d after rainfall, and the third sampling point was just before plant establishment (Fig. S1).



At day 48 after initiating the experiment, each mesocosm was planted with 10 seedlings, one of each exotic and native plant species that co-occur in riparian areas in northwestern Europe (Table S1). Seedlings were planted in a random order with 5 cm distance and grown for 69 d. After 21 d of plant growth, half the planted mesocosms received a 14-d drought period until the soil moisture was 3.6% wt/wt, after which soils were rewetted up to 20% wt/wt (Fig. S1). Mesocosms were placed randomly in a greenhouse and received 16 h daylight per 24 h. At day 117, roots and shoots were harvested, sorted to species and dried to constant weight at 70 °C. Seven of the 64 mesocosms, which were distributed across the eight treatments, were excluded from further analyses because they did not contain the intended plant species owing to seed pollution. Fig. 1 shows the number of mesocosms per treatment.

**Experiment 2: Effect of Biotic Legacy on Plant Species.** For this experiment, we established the same plant communities as in experiment 1 in 56 mesocosms and grew the plants for 70 d. The mesocosms contained the sterilized bulk soil (see above) that was mixed with either living or sterilized soil inocula collected at day 46 of experiment 1. The inocula originated from seven replicates of the four drought/rainfall soil treatments (Fig. S1A). The inoculum from each pot was kept separate and split into two halves (equivalent to 325 g dry soil each). One half was sterilized twice at 120 °C for 1 h in an autoclave with a 48-h time interval to kill all soil biota present in the inocula (41). We used autoclaving instead of  $\gamma$ -irradiation, because of accessibility limitations of the irradiation facilities for small soil samples. The other half of the soil was kept untreated and served as living inoculum. The sterilized and living

soil inocula were mixed with the sterilized bulk soil according to a 1:8 mixture of inoculum and bulk soil. The soil mixture was placed in 3-L mesocosms and planted after 3 d. It is well known that soil sterilization increases soil nutrient availability (43, 44). Considering that we added one part of inoculum to 8 parts (wt/wt) of sterilized background soil, and given that we worked with a nutrient-poor sandy soil, effects of nutrient flush differences between sterilized and living inocula will have been overwhelmed by nutrients in the sterilized background soils. This will have increased the likelihood that differences among treatments will have been caused by changes in soil biota. We established the same plant communities as in experiment 1 in all mesocosms and grew, harvested, and weighed plants as in experiment 1.

**Data Analysis.** Data were analyzed using SPSS 17 (IBM) and CANOCO 2.55 (Ter Braak and Šmilauer). Details are presented in *SI Materials and Methods*.

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